



34 **Abstract**

35 A large fraction of the world grasslands and savannas are undergoing a rapid shift from  
36 herbaceous to woody-plant dominance. This land-cover change is expected to lead to a loss in  
37 livestock production, but the impacts of woody-plant encroachment on this crucial ecosystem  
38 service have not been assessed. We evaluated how tree cover has affected livestock production at  
39 large scales in rangelands of contrasting socio-economic characteristics in the U.S. and  
40 Argentina. Our models indicated that in areas of high productivity, a 1% increase in tree cover  
41 resulted in a reduction in livestock production ranging from 0.6 to 1.6 reproductive cows per  
42  $\text{km}^2$ . Mean livestock production in the U.S. is  $27 \text{ Rc km}^{-2}$ , so a 1% increase in tree cover results  
43 in 2.5% decrease in mean livestock production. This effect is large considering that woody-plant  
44 cover is increasing at 0.5-2% per year. On the contrary, in areas of low productivity, increased  
45 tree cover had a positive effect on livestock production. Our results also show that ecological  
46 factors account for a larger fraction of livestock production variability in Argentinean than in  
47 U.S. rangelands. Differences in the relative importance of ecological versus non-ecological  
48 drivers of livestock production in Argentina and the U.S. suggest that the valuation of ecosystem  
49 services between these two rangelands might be different. Current management strategies in  
50 Argentina are likely designed to maximize livestock production whereas land managers in the  
51 U.S. may be optimizing multiple ecosystem services, including conservation or recreation,  
52 alongside livestock production.

53

54 **Significance Statement**

55 Grasslands all over the world are undergoing a rapid shift from a regime dominated by  
56 herbaceous plants to one dominated by woody plants, a phenomenon known as woody-plant  
57 encroachment. The impact of this global phenomenon on livestock production, the main  
58 ecosystem service provided by grasslands, remains largely unexplored. We quantified, for the  
59 first time, the impact of woody-plant encroachment on livestock production at a large scale,  
60 finding a reduction of between 0.6 and 1.6 reproductive cows per km<sup>2</sup> for each 1% increase in  
61 tree cover. By comparing the largest rangelands of the Americas (U.S. and Argentina), we also  
62 showed how the impact of woody-plant encroachment is mediated by social-economic factors.  
63 Our manuscript represents a significant advance in our understanding of grasslands as complex  
64 social-ecological systems.

65

66

67 **Introduction**

68 Grasslands, shrublands, and savannas, collectively ‘rangelands’, constitute ca. 50% of the Earth’s  
69 land surface (1). Although characterized by low yet highly variable annual rainfall, these areas  
70 provide 30-35% of terrestrial net primary productivity (2), contain >30% of the world’s human  
71 population, and support the majority of the world’s livestock production (3, 4). Besides  
72 livestock production, rangelands also provide a variety of other ecosystem services, including  
73 fiber production, carbon sequestration, maintenance of the genetic library (conservation), and  
74 recreation (5).

75 One of the most striking land-cover changes in rangelands worldwide over the past 150  
76 years has been the proliferation of trees and shrubs at the expense of perennial grasses (6). In the  
77 U.S., non-forest lands undergoing woody-plant encroachment are now estimated to cover up to  
78 335 million ha, i.e. 40% of the coterminous U.S., (7) and the increase in woody cover ranges  
79 from 0.5 to 2% per year (8). The causes of this vegetation change are debated and the main  
80 potential drivers include intensification of livestock grazing, changes in climate and fire regimes,  
81 the introduction of non-native woody species, and declines (natural and human-induced) in the  
82 abundance of browsing animals (9-12). Historical increases in atmospheric nitrogen deposition  
83 and atmospheric carbon dioxide concentration have also been suggested to play a role (10, 11).

84 Woody-plant encroachment has long been of concern to a broad range of stakeholders,  
85 from pastoralists to ranchers, because of the expected negative impact on livestock production  
86 (13). In response, brush management has been widely used to reduce the cover of encroaching  
87 woody-vegetation on both public and private lands. For example, the U.S. Natural Resources  
88 Conservation Service spent \$127M in brush management programs in the period 2005-2009,  
89 implemented on more than 1 million ha of rangeland (14). Despite claims about impact of  
90 woody-plant encroachment on livestock production and the large amounts of federal, state, and  
91 private spending on brush management, the impact of woody-plant encroachment on livestock  
92 production has seldom been quantified (15). Here, our objectives are i) to quantify how woody-  
93 plant encroachment affects livestock production at large spatial scales, and ii) to assess how this  
94 impact is modified under different ecological and social-economic conditions.

95 We developed a general framework in which livestock production depends on net  
96 primary production, woody-plant cover, and other non-biological determinants. Net primary  
97 production sets the total amount of biomass and energy that is available to herbivores (16). The

98 most common view on woody-plant encroachment is that encroachment diverts herbaceous  
99 productivity, on which cattle feed, to unpalatable woody-plant productivity, thus reducing  
100 potential energy intake (17-19). Thus overall, primary production and woody-plant  
101 encroachment jointly determine the livestock carrying capacity of an ecosystem.

102 In natural ecosystems from forests to deserts, there is a tight correlation among primary  
103 productivity and secondary productivity and animal biomass (16). Social and economic factors  
104 determine how close current livestock stocking is to the carrying capacity of the site, which is  
105 determined by NPP. Oesterheld et al. (20) assessed the relationship between net primary  
106 productivity and livestock production in managed rangelands in Argentina, where management  
107 focuses on food production, and found that the link between primary and secondary productivity  
108 was even tighter than in natural ecosystems. Management practices such as providing water and  
109 minerals, regulating animal distribution, and reducing parasitism, predation and diseases,  
110 resulted in stocking rates that were closely associated with net primary productivity.

111 We expect that in advanced industrial societies, where the production of goods (e.g., food  
112 by means of agriculture and ranching) plays a secondary role in the economy (21), landscapes  
113 will be managed to maximize multiple ecosystem services, and thus livestock production might  
114 be less driven by ecological drivers. Ecological factors, including net primary productivity and  
115 woody-plant cover, determine potential stocking rate but actual stocking rate is modulated by  
116 manager's decisions (22). In some cases, land managers overstock rangelands leading to  
117 degradation and desertification (23) while in other cases managers understock. The latter results  
118 from pursuing optimization of multiple ecosystem services of which food production is only one.  
119 Rangelands managed for multiple purposes and ecosystem services (24) seek provisioning of  
120 food, fiber, firewood, carbon sequestration, conservation or recreation.

121 Our hypotheses are i) that overall livestock production decreases with woody-plant  
122 encroachment, ii) the effect of woody-plant encroachment on livestock production is modulated  
123 by NPP, with a larger negative impact of woody-plant encroachment in those areas with higher  
124 net primary productivity, and iii) the role of ecological drivers (net primary productivity and tree  
125 cover) on livestock production is larger in regions where the demand for ecosystem services is  
126 concentrated exclusively on food production.

127 The scarcity of studies attempting to quantify the impact of woody-plant encroachment  
128 on livestock production reflects the enormous difficulties of addressing this issue by means of

129 conventional field approaches. An experimental approach necessitates monitoring the change in  
130 livestock production in a number of locations during the encroachment process, a process that  
131 might take decades (11). Our approach has been to explore how current rangeland livestock  
132 production varies at a regional scale along sites with different net primary productivity and  
133 woody cover. We thus assessed the consequences of the process of woody-plant encroachment  
134 by evaluating the relationship between tree-cover and livestock production at a given point in  
135 time across multiple locations. This approach of swapping time for space has been used to  
136 predict future trajectories of species in an ecological succession (25), and more recently, the  
137 expected change of organisms ranging from microbes (26) to trees (27) under a changing  
138 climate. We are aware of the limitations of the approach mostly associated with the existence of  
139 lags that result in different models through space and time (28). Given the limitations of  
140 alternative options and the urgency of the problem, we consider our approach to be promising.

141 To test our hypotheses, we collected information about woody-plant cover and primary  
142 productivity from remote sensing sources and about livestock production from agricultural  
143 census data. Woody-plant encroachment occurs when there is an increase in the abundance of  
144 trees or shrubs. The type of woody component depends on mean annual precipitation with arid  
145 systems being invaded by shrubs and mesic ecosystems being invaded by trees. In our study  
146 areas, the transition between shrub and tree domains occurs approximately at 600 mm of annual  
147 precipitation (Fig S2). In the present work, we focused on encroachment of trees (i.e., areas >600  
148 mm) because current remote sensing tools assess tree cover with accuracy, but they do not  
149 adequately estimate shrub cover (29) and thus our approach is not feasible in shrublands. We  
150 aggregated data at the county level and combined remote sensing and census data in a model that  
151 yields estimates of the impact of woody-plant cover on livestock production at large scales. To  
152 account for the effects of socio-economic factors, we quantified the impact of tree cover on  
153 livestock production in two regions of the world that have extraordinary environmental similarity  
154 but have contrasting socio-economic characteristics (30, 31). The two regions are the U.S.  
155 Central Grassland Region and the Argentinean Central Grassland. Both share similar temperature  
156 and precipitation gradients, yielding vegetation types that are remarkably similar (31) (Fig. 1).  
157 These environmental similarities contrast with large socio-economic differences in the rural  
158 sector, and specifically regarding livestock production (Supplementary Information, Fig. S1).  
159 During the last decades in the U.S., there has been a reduction of people making a living from

160 agriculture (40% reduction since 80s) and a negative trend in the number of cattle in the region  
161 (22% reduction since the 70s). At present, a large proportion of stakeholders in the U.S. are not  
162 full-time ranchers but maintain livestock production as a source of secondary income or for  
163 cultural or recreational reasons (USDA economic service; [www.ers.usda.gov](http://www.ers.usda.gov)). In Argentina,  
164 although the relative importance of ranching has decreased due to the expansion of crop  
165 products, especially soybean, the reduction in the number of cattle has been much smaller (4%  
166 reduction since the 70s, Fig. S1); and beef is still the agricultural commodity with the largest  
167 output value (28% of the total agricultural production 2005-2007) (32). As a result, we expected  
168 stocking rates in Argentina to be closer to the NPP-derived carrying capacity of the system, and  
169 thus more tightly driven by ecological factors, than in the U.S. (20).

170

## 171 **Results and Discussion**

172 In both the U.S. and Argentina, livestock production shows a W to E gradient of increasing  
173 reproductive density. The maximum value in the U.S. is 66 reproductive cows (Rc) per km<sup>2</sup> in  
174 the eastern part of the region. In Argentina, this gradient is more apparent than in the U.S.,  
175 reaching maximum values of 43 Rc km<sup>-2</sup> (Fig. 2). This directional gradient is the same for NPP  
176 and tree cover gradients in both regions, following mean annual precipitation gradients (Fig. 1).

177 In accordance with our first hypothesis, woody-plant encroachment in both rangelands  
178 had a negative impact on livestock production. An increase of 1% in tree cover resulted in an  
179 overall decrease in livestock production ranging from 0.6 to 1.6 Rc km<sup>-2</sup> (Fig. 3, Table 1). In the  
180 U.S., an increase in tree cover of 1% decreased livestock production by 0.57 Rc km<sup>-2</sup>. Mean  
181 livestock production in the U.S. is 27 Rc km<sup>-2</sup>, so a 1% increase in tree cover results in 2.5%  
182 decrease in mean livestock production of the region. In NPP units, a 1% increase in tree cover  
183 had the same impact on livestock production as an NPP decrease of 41 g C m<sup>-2</sup> y<sup>-1</sup>. The  
184 magnitude of the impact can be gauged when taking into account that, in North America, the  
185 increase of woody cover ranges from 0.5 to 2% per year (8).

186 As in our second hypothesis, in Argentina, a significant interaction between NPP and tree  
187 cover as drivers of livestock production exists, although we did not find this interaction when  
188 evaluating the U.S. data (Fig. 3, Table 1). At high productivity values (900 g C m<sup>-2</sup> y<sup>-1</sup>), an  
189 increase of 1% tree cover decreased livestock production by 1.6 Rc km<sup>-2</sup>, relative to livestock  
190 production ranging between 1 and 43 Rc km<sup>-2</sup>. However, at productivity values of less than 365

191  $\text{g C m}^{-2} \text{ y}^{-1}$ , tree cover enhanced livestock production. In low productivity ( $300 \text{ g C m}^{-2} \text{ y}^{-1}$ ) areas  
192 in Argentina, an increase in tree cover of 1% increased livestock production by  $0.24 \text{ Rc km}^{-2}$ .  
193 This result contradicts current understanding of the impact of woody-plant encroachment, which  
194 is thought to have a negative impact on livestock production (6, 17-19, 33). Note that the lower  
195 limit of NPP in our study area in the U.S. occurs above  $365 \text{ g C m}^{-2} \text{ y}^{-1}$ , obscuring a possible  
196 positive effect of tree cover on livestock production at low productivity values. Potential  
197 explanations of this positive effect of woody-plant encroachment on livestock production at low  
198 productivity values may be found in factors other than the amount of food available for livestock  
199 production. For example, most of the areas of low productivity in our study area are associated  
200 with low precipitation and high temperature (Fig. 1). In these areas, tree cover might provide  
201 shelter and shade or overall near-ground temperatures, decreasing animal respiration costs (34) .

202 Our results showed that the effect of NPP and tree cover on productivity was larger in  
203 Argentina than in the U.S. ( $R^2 = 50\%$  and  $24\%$  respectively, Table 1), indicating a strong  
204 difference between the two study areas in the importance of the drivers of livestock production.  
205 This aligns with our third hypothesis, that the role of ecological drivers (net primary productivity  
206 and tree cover) on livestock production would be larger in regions where the demand for  
207 ecosystem services is concentrated exclusively on food production. The effect of tree cover on  
208 livestock production relative to the effect of net primary productivity on livestock productivity  
209 was similar in the two study regions, with the explanatory power of NPP being five times larger  
210 than that of tree cover (U.S.:  $R^2_{\text{NPP}} = 20\%$  and  $R^2_{\text{TC}} = 4\%$ ; Argentina:  $R^2_{\text{NPP}} = 42\%$  and  $R^2_{\text{TC}} = 8\%$ ,  
211 being  $R^2_{\text{NPP}}$  and  $R^2_{\text{TC}}$  the percentage of variance accounted for by net primary productivity and  
212 tree cover) (Fig. 4). The similarity in the relative importance of NPP and tree cover indicates  
213 that, despite the difference in socio-economic differences, the underlying ecological mechanisms  
214 driving livestock production are similar.

215 Differences in the relative importance of ecological vs. non-ecological (social) drivers on  
216 livestock production in Argentina and the U.S. suggest that the value of the various ecosystem  
217 services provided by rangelands may be different in these two regions. Rangelands produce a  
218 variety of ecosystem services including food and fiber production, carbon sequestration,  
219 maintenance of the genetic library (conservation), and recreation (5). Current management  
220 strategies in Argentina are likely to be designed to maximize a single ecosystem service  
221 (livestock production). On the contrary, land managers in the U.S. appear to be optimizing

222 multiple ecosystem services, including conservation or recreation alongside livestock production.  
223 Therefore, it is important to measure the effects of woody-plant encroachment on the entire  
224 portfolio of ecosystem services that are provided by rangelands. Most changes in ecosystem  
225 services due to woody-plant encroachment remain unclear and have been identified only in a  
226 qualitative fashion (but see (33)). Future quantitative studies taking into account multiple  
227 ecosystems services are needed in order to assist in decision making whether to implement or not  
228 brush management actions. Livestock production is currently one of the most important  
229 ecosystem service provided by rangelands but the development trajectory highlighted by the  
230 differences between Argentina and the U.S. point out that other ecosystem services will likely  
231 become increasingly important as economies undergo a transition from the production of goods  
232 to the provision of services.

233 Our study demonstrates that livestock production is part of an integrated socio-ecological  
234 system where ecological and social-economic drivers interact along gradients of climate and  
235 economic development (22). In high productivity regions, woody-plant cover negatively affects  
236 livestock production mainly through reductions in forage availability. The negative effect of  
237 woody plants on forage availability is overwhelmed in low productivity regions by the positive  
238 effects of woody cover that may be linked to the amelioration temperature, a possible linkage  
239 that requires examination. As economic development increases the demand for ecosystem  
240 services from rangelands becomes more diversified. In least developed regions, food and fiber  
241 dominate the demand for ecosystem services. On the contrary, in developed regions there are  
242 multiple demands from rangelands beyond food production that include conservation, carbon  
243 sequestration, water supply and recreation. As development increases and demand diversifies the  
244 importance of ecological drivers decreases while that of social-economic factors increases. The  
245 future of woody-plant encroachment and its consequences on ecosystem services will be  
246 modulated by changing climate and social and economic conditions.

247

## 248 **Methods Summary**

### 249 **Study areas**

250 We modeled the impact of woody plant encroachment on livestock production at a county  
251 resolution for both U.S. and Argentinean rangelands (Fig. 1). Both areas share a similar  
252 latitudinal temperature gradient and a longitudinal precipitation gradient, with precipitation

253 increasing from W to E. These similar climatic patterns yield vegetation types that are  
254 remarkably similar (31). These similarities contrast with large social-economic differences (see  
255 Introduction and Figure S1), which make them a perfect study system to address the impact of  
256 woody-plant encroachment on livestock production at a regional scale and the variation of this  
257 impact between different socio-economic regions.

258 The U.S. and Argentinean rangelands constitute, together with the Brazilian Cerrado, the  
259 two main rangelands of the Americas (35). Here, we used rangelands in a very broad sense; our  
260 two study areas comprise the transition between the desert and the forest biomes. We defined our  
261 study areas in the U.S. and Argentina as those encompassing the prairie, savanna, and temperate  
262 and subtropical desert and steppes divisions and regimen mountain divisions, according to  
263 Bailey's ecoregions (1). Within those areas, we excluded those counties with mean annual  
264 precipitation values below 600 mm, thus focusing on the tree dominion (Fig S2) and excluding  
265 woody cover due to shrubs. The resulting areas in the U.S. and Argentina had the same  
266 precipitation lower limit (600mm) but differed in their upper limit (U.S.=1260 mm, Argentina=  
267 2270 mm). In order to make the analysis of both areas fully comparable we limited the upper  
268 precipitation limit of Argentina to that of the U.S. (i.e., 1260 mm). Taking into account also  
269 those counties excluded due to low representation of non-crop lands (see below), the resulting  
270 number of sampling units (i.e., counties in the U.S., departments in Argentina) was 242 for the  
271 U.S. and 125 for Argentina.

272

### 273 **Livestock production data**

274 Data on livestock production were obtained from the USDA Census Database  
275 ([www.agcensus.usda.gov](http://www.agcensus.usda.gov)) and Argentinean Food and Agriculture Administration (SENASA;  
276 <http://www.senasa.gov.ar>) (Fig. 2). In both cases, we used the last available livestock data (2007  
277 for the U.S., and 2010 for Argentina). We focused on cattle, which is the main livestock type in  
278 both areas. For comparability, we used the number of reproductive animals, a metric present in  
279 both data bases. This metric corresponded to the class 'Cows incl. calves' in the USDA Census  
280 data and to the class 'Cows' in the SENASA database (range: 1.5-66.4 and 0.5-43.2 animals per  
281 km<sup>2</sup> for the U.S. and Argentina respectively). In the U.S. we subtracted the number of cows on  
282 feedlots, also available in the U.S. Census Database, from the total number of cows.

283

## 284 **Environmental data**

285 Net primary productivity, tree cover, and land uses per county were quantified by using  
286 Moderate-resolution Imaging Spectroradiometer (MODIS) products (Fig. 2). All environmental  
287 variables were characterized by the mean annual values of the year of the livestock data (2007  
288 for the U.S., and 2010 for Argentina) and the four previous years. The value of the Net primary  
289 productivity was assessed using the Photosynthesis and Net Primary Productivity algorithm  
290 MOD17A3 (36). Here, production is determined by first computing a daily net photosynthesis  
291 value which is then composited over an 8-day interval of observations over a year, to produce a  
292 net primary productivity measure. Tree cover was assessed by means of MODIS Vegetation  
293 Continuous Fields product MOD44B (29). This product represents Earth's terrestrial surface as a  
294 proportion of three surface cover components: percent tree cover, percent non-tree cover, and  
295 percent bare ground. Land uses were assessed by the MODIS product MCD12Q1 (37). This  
296 land-use remote sensing data allowed us to exclude crops and urban areas in our analysis, and  
297 thus to obtain a more accurate measure of the net primary productivity available for livestock  
298 consumption per county. Additionally, in order to remove those counties with low sampling size,  
299 we also excluded from our analyses those counties with less than 1000 km<sup>2</sup> or 25% of  
300 rangelands.

301 Mean annual precipitation values were obtained from Earth observations and climatic  
302 models. Specifically, annual precipitation values for the study periods in Argentina were  
303 obtained from the Tropical Measuring Mission (TRMM; [www.trmm.gsfc.nasa.gov](http://www.trmm.gsfc.nasa.gov)) at a 0.25°  
304 resolution. In the U.S., annual climatic data at a 2.5' resolution were obtained from the PRISM  
305 Climate Group (Oregon State University; [www.prism.oregonstate.edu](http://www.prism.oregonstate.edu)).

306

## 307 **Hypotheses testing**

308 Our first two hypotheses describe the impact of net primary productivity and tree cover on  
309 livestock production and were tested by means of the model  $LP = \beta_0 + \beta_1 * NPP + \beta_2 * TC +$   
310  $\beta_3 * NPP * TC$ , where LP is livestock production, NPP is net primary productivity and TC is tree  
311 cover. The sign and significance of  $\beta_2$  and  $\beta_3$  in the models fitted for the two study areas (U.S.  
312 and Argentinean rangelands) tested first and second hypotheses.

313 The third hypothesis, that states that the role of ecological drivers (net primary productivity and  
314 tree cover) on woody-plant encroachment on livestock production would be larger in regions

315 where the demand for ecosystem services is concentrated exclusively on food production, was  
316 tested by means of examining model results in the U.S. and Argentinean rangelands separately.  
317 In particular, we examined the explained variance of the model in each country. The relative  
318 explanatory power of NPP and tree cover was assessed by means of a variance partitioning  
319 analysis (38, 39), which allowed us to break down the total explained variance in four fractions:  
320 pure effects of NPP, pure effects of TC (i.e., variance exclusively explained by NPP or TC), join  
321 effect of NPP and TC (i.e., variance explained simultaneously by NPP and TC) and the effect of  
322 the synergistic interaction between the two drivers (variance explained by NPP\*TC).

323 The model  $LP = \beta_0 + \beta_1*NPP + \beta_2*TC + \beta_3*NPP*TC$  was fitted with three candidate sets  
324 of variables describing NPP and TC considering one, three or five years of previous information:  
325 a) variables describing NPP and TC values for the year of census (2007 for the US and 2010 for  
326 Argentina), b) variables describing the average NPP and TC values of the year of the census and  
327 the two previous years and c) the average NPP and TC values of the year of the census and the  
328 four previous years. For both the U.S. and Argentina, the three candidate set of variables yielded  
329 very similar patterns, although the models with largest values of explained variance, and thus  
330 those presented here, were those with independent variables describing NPP and TC the year  
331 before the livestock census data.

332

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### 341 **REFERENCES**

- 342 1. Bailey RG Ropes L (1998) *Ecoregions: the ecosystem geography of the oceans and*  
343 *continents* (Springer New York).
- 344 2. Field CB, Behrenfeld MJ, Randerson JT, Falkowski P (1998) Primary production of the  
345 biosphere: integrating terrestrial and oceanic components. *Science* 281(5374):237-240.

- 346 3. Safriel U Adeel Z (2005) Dryland systems. *Ecosystems and human well-being, current*  
347 *state and trends*, eds Hassan R, Scholes R, & Ash N (Island Press, Washington), Vol 1,  
348 pp 625-658.
- 349 4. Reynolds JF, et al. (2007) Global desertification: building a science for dryland  
350 development. *Science* 316(5826):847-851.
- 351 5. Sala O Paruelo J (1997) Ecosystem services in grasslands. *Nature's services: Societal*  
352 *dependence on natural ecosystems*, ed Daily GC (Island Press, Washington, D.C.), pp  
353 237-251.
- 354 6. Archer SR (2010) Rangeland conservation and shrub encroachment: new perspectives on  
355 an old problem. *Wild rangelands: Conserving wildlife while maintaining livestock in*  
356 *semi-arid ecosystems*, eds du Toit JT, Kock R, & C DJ (John Wiley & Sons, Ltd,  
357 Chichester, UK), pp 53-97.
- 358 7. Pacala SW, et al. (2001) Consistent land-and atmosphere-based US carbon sink  
359 estimates. *Science* 292(5525):2316-2320.
- 360 8. Barger NN, et al. (2011) Woody plant proliferation in North American drylands: a  
361 synthesis of impacts on ecosystem carbon balance. *Journal of Geophysical Research*  
362 116(G00K07).
- 363 9. Archer S (1995) Tree-grass dynamics in a Prosopis-thornscrub savanna parkland:  
364 reconstructing the past and predicting the future. *Ecoscience* 2(1):83-99.
- 365 10. Van Auken OW (2000) Shrub invasions of North American semiarid grasslands. *Annual*  
366 *Review of Ecology and Systematics* 31:197-215.
- 367 11. Briggs JM, et al. (2005) An ecosystem in transition: causes and consequences of the  
368 conversion of mesic grassland to shrubland. *BioScience* 55(3):243-254.
- 369 12. Naito AT Cairns DM (2011) Patterns and processes of global shrub expansion. *Progress*  
370 *in Physical Geography* 35(4):423-442.
- 371 13. Scholes R Archer S (1997) Tree-grass interactions in savannas. *Annual Review of*  
372 *Ecology and Systematics* 28:517-544.
- 373 14. Tanaka JA, Torell LA, Brunson MW, Briske DD (2011) A social and economic  
374 assessment of rangeland conservation practices. *Conservation benefits of rangeland*  
375 *practices: assessment, recommendations, and knowledge gaps*, ed Briske DE (US  
376 Department of Agriculture, Natural Resources Conservation Service, Washington, DC),  
377 pp 371-422.
- 378 15. Dalle G, Maass BL, Isselstein J (2006) Encroachment of woody plants and its impact on  
379 pastoral livestock production in the Borana lowlands, southern Oromia, Ethiopia. *African*  
380 *Journal of Ecology* 44(2):237-246.
- 381 16. McNaughton SJ, Oesterheld M, Frank DA, Williams KJ (1989) Ecosystem-level patterns  
382 of primary productivity and herbivory in terrestrial habitats. *Nature* 341(6238):142-144.
- 383 17. Ethridge D, Dahl B, Sosebee R (1984) Economic evaluation of chemical mesquite control  
384 using 2, 4, 5-T. *Journal of Range Management* 37(2):152-156.
- 385 18. Hedrick DW, Hyder DN, Sneva FA, Poulton CE (1966) Ecological response of  
386 sagebrush-grass range in central Oregon to mechanical and chemical removal of  
387 *Artemisia*. *Ecology* 47(3):432-439.
- 388 19. Hyder DN Sneva FA (1956) Herbage response to sagebrush spraying. *Journal of Range*  
389 *Management* 9(1):34-38.
- 390 20. Oesterheld M, Sala OE, McNaughton SJ (1992) Effect of animal husbandry on herbivore-  
391 carrying capacity at a regional scale. *Nature* 356(6366):234-236.

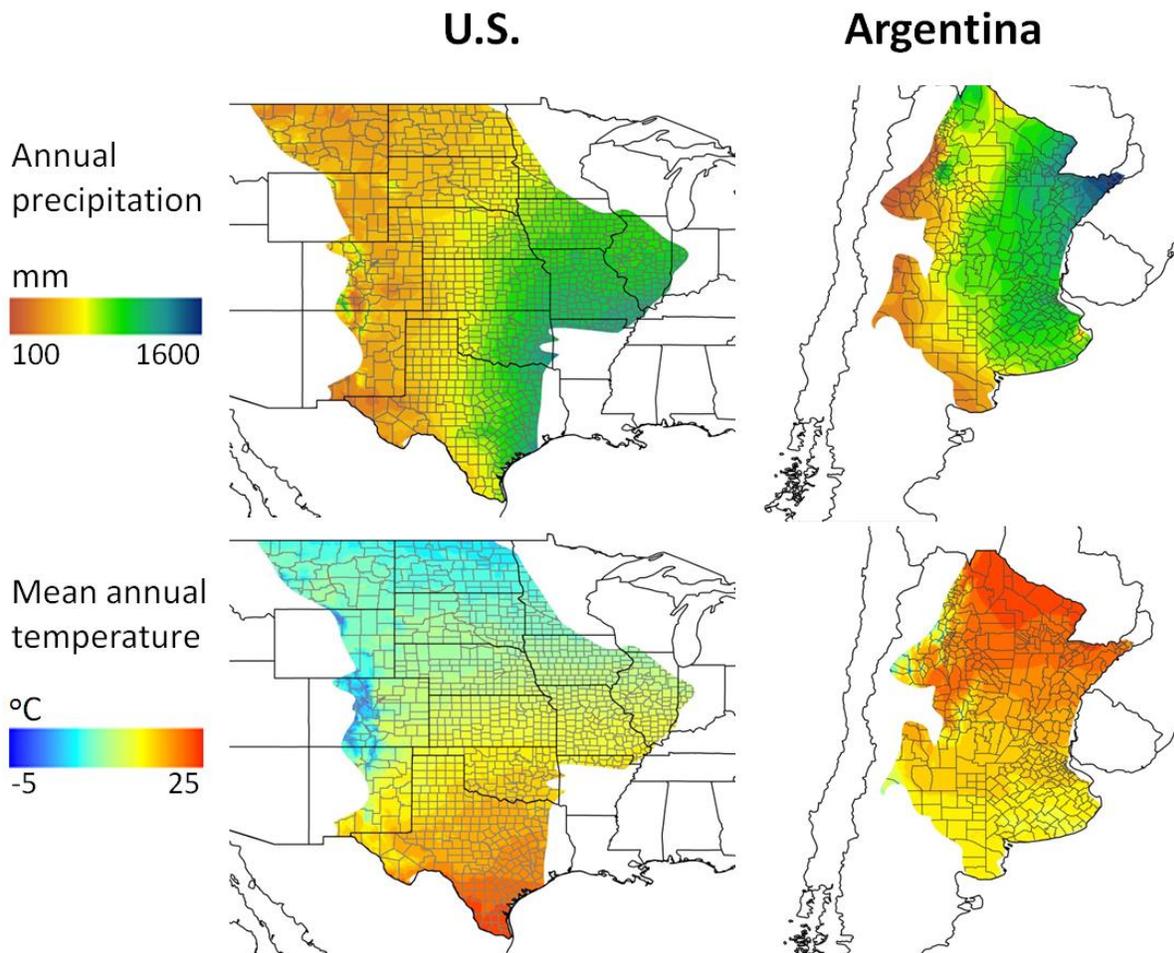
- 392 21. Bell D (1973) *The coming of post-industrial society: a venture in social forecasting.*  
393 (Basic Books, New York), p 507.
- 394 22. Díaz S, et al. (2011) Linking functional diversity and social actor strategies in a  
395 framework for interdisciplinary analysis of nature's benefits to society. *Proceedings of*  
396 *the National Academy of Sciences* 108(3):895-902.
- 397 23. Geist HJ Lambin EF (2004) Dynamic causal patterns of desertification. *BioScience*  
398 54(9):817-829.
- 399 24. Pratt D Gwynne MD (1977) *Rangeland management and ecology in East Africa* (Hodder  
400 and Stoughton).
- 401 25. Clements FE (1916) *Plant succession: an analysis of the development of vegetation*  
402 (Carnegie Institution of Washington, Washington, DC) p 512.
- 403 26. Garcia-Pichel F, Loza V, Marusenko Y, Mateo P, Potrafka RM (2013) Temperature  
404 Drives the Continental-Scale Distribution of Key Microbes in Topsoil Communities.  
405 *Science* 340(6140):1574-1577.
- 406 27. Thuiller W, Lavorel S, Araújo MB, Sykes MT, Prentice IC (2005) Climate change threats  
407 to plant diversity in Europe. *Proceedings of the National Academy of Sciences of the*  
408 *United States of America* 102(23):8245-8250.
- 409 28. Sala OE, Gherardi LA, Reichmann L, Jobbágy E, Peters D (2012) Legacies of  
410 precipitation fluctuations on primary production: theory and data synthesis. *Philosophical*  
411 *Transactions of the Royal Society B: Biological Sciences* 367(1606):3135-3144.
- 412 29. Hansen MC, et al. (2003) Global percent tree cover at a spatial resolution of 500 meters:  
413 First results of the MODIS vegetation continuous fields algorithm. *Earth Interactions*  
414 7(10):1-15.
- 415 30. Paruelo JM, et al. (1995) Regional climatic similarities in the temperate zones of North  
416 and South America. *Journal of Biogeography* 22(4/5):915-925.
- 417 31. Paruelo JM, Jobbágy EG, Sala OE, Lauenroth WK, Burke IC (1998) Functional and  
418 structural convergence of temperate grassland and shrubland ecosystems. *Ecological*  
419 *Applications* 8(1):194-206.
- 420 32. Lence SH (2010) The agricultural sector in Argentina: major trends and recent  
421 developments. *The shifting patterns of agricultural production and productivity*  
422 *worldwide*, eds Alston JM, Babcock BA, & Pardey PG (The Midwest Agribusiness Trade  
423 Research and Information Center, Iowa State University, Ames, Iowa), pp 409-448.
- 424 33. Eldridge DJ, et al. (2011) Impacts of shrub encroachment on ecosystem structure and  
425 functioning: towards a global synthesis. *Ecology Letters* 14(7):709-722.
- 426 34. Tucker CB, Rogers AR, Schütz KE (2008) Effect of solar radiation on dairy cattle  
427 behaviour, use of shade and body temperature in a pasture-based system. *Applied Animal*  
428 *Behaviour Science* 109(2-4):141-154.
- 429 35. Wint W Robinson TP (2007) *Gridded livestock of the world 2007* (Food and Agriculture  
430 Organization of the United Nations Rome).
- 431 36. Heinsch FA, et al. (2003) User's Guide GPP and NPP (MOD17A2/A3) Products NASA  
432 MODIS Land Algorithm. pp 1-57.
- 433 37. Friedl MA, et al. (2010) MODIS Collection 5 global land cover: Algorithm refinements  
434 and characterization of new datasets. *Remote Sensing of Environment* 114(1):168-182.
- 435 38. Borcard D, Legendre P, Drapeau P (1992) Partialling out the spatial component of  
436 ecological variation. *Ecology* 73(3):1045-1055.

- 437 39. Cushman SA McGarigal K (2002) Hierarchical, multi-scale decomposition of species-  
438 environment relationships. *Landscape Ecology* 17(7):637-646.  
439  
440

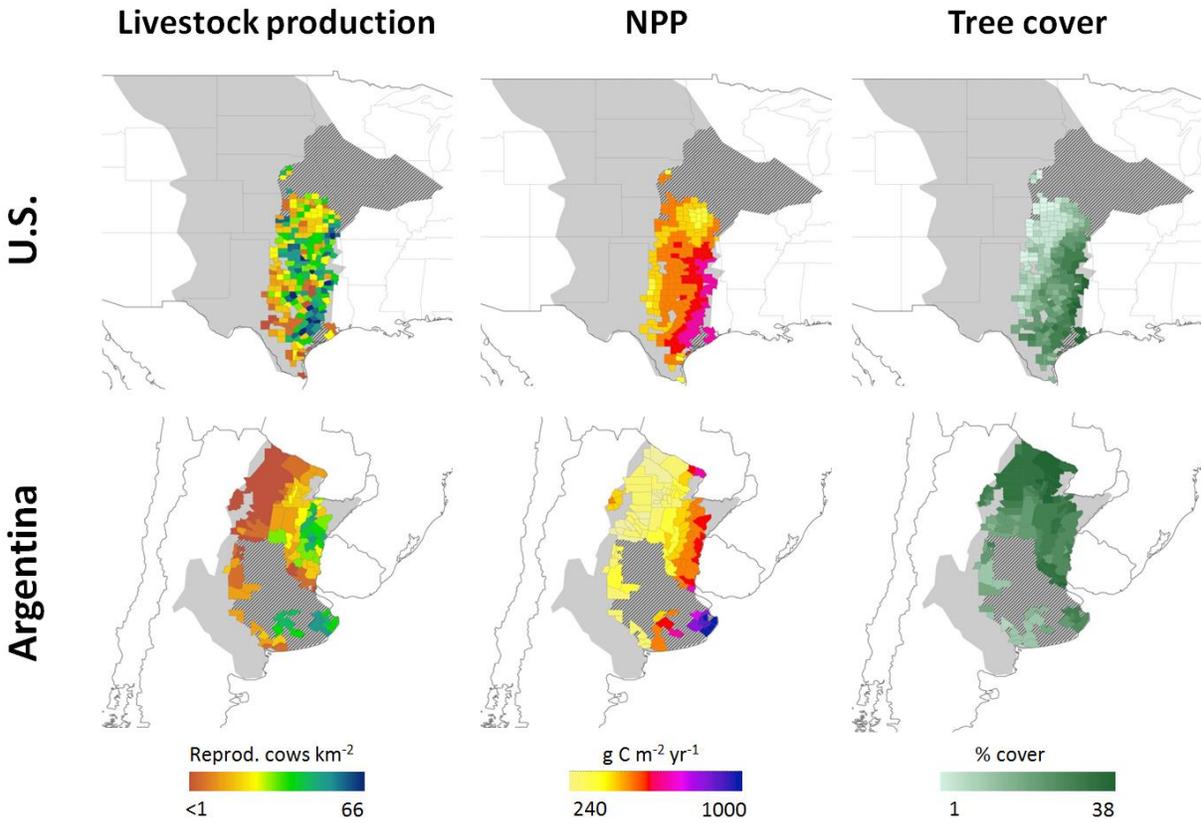
**Table 1.** Models assessing the effect of ecological drivers (net primary productivity and tree cover) on livestock production in the U.S. and Argentinean rangelands.  $R^2$  = % of explained variance. N.s. = non-significant effect (not included in the final model).

	U.S.		Argentina	
	Estimate	p-value	Estimate	p-value
Intercept	-40.8044	0.8424	-22.75	0.6015
NPP	0.133	<0.0001	0.09796	<0.0001
Tree cover	-0.5754	0.0005	1.1360	0.0006
NPP*Tree cover	-	n.s.	-0.003	0.0001
$R^2$	24.01		50.26	

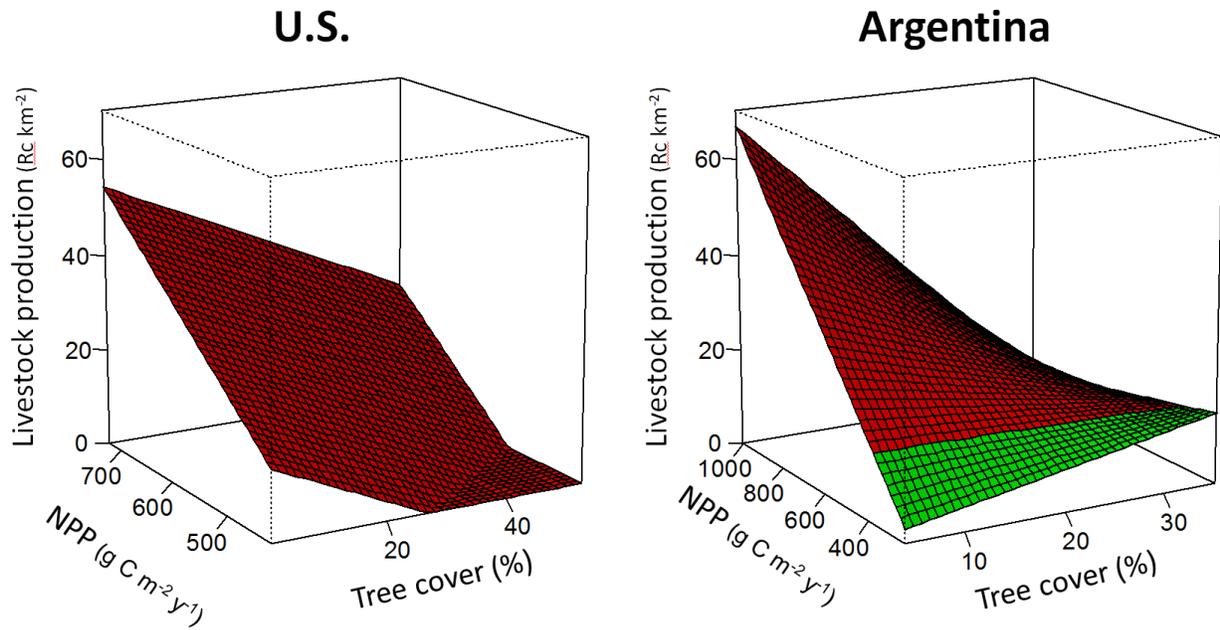
**Figure 1.** Main environmental gradients (mean annual precipitation and mean annual temperature) in the U.S. and Argentinean rangelands. Rangelands are defined in this work as those areas encompassing the prairie, savanna, and temperate and subtropical desert, steppes and mountain divisions, according to Bailey’s ecoregions (1). Within these areas our work focused on those counties with mean annual precipitation values between 600 and 1260 (see Methods Section and Fig. 2). For both areas, national (bold lines) and county (thin lines) borders are drawn. In the US state borders are also drawn (bold lines).



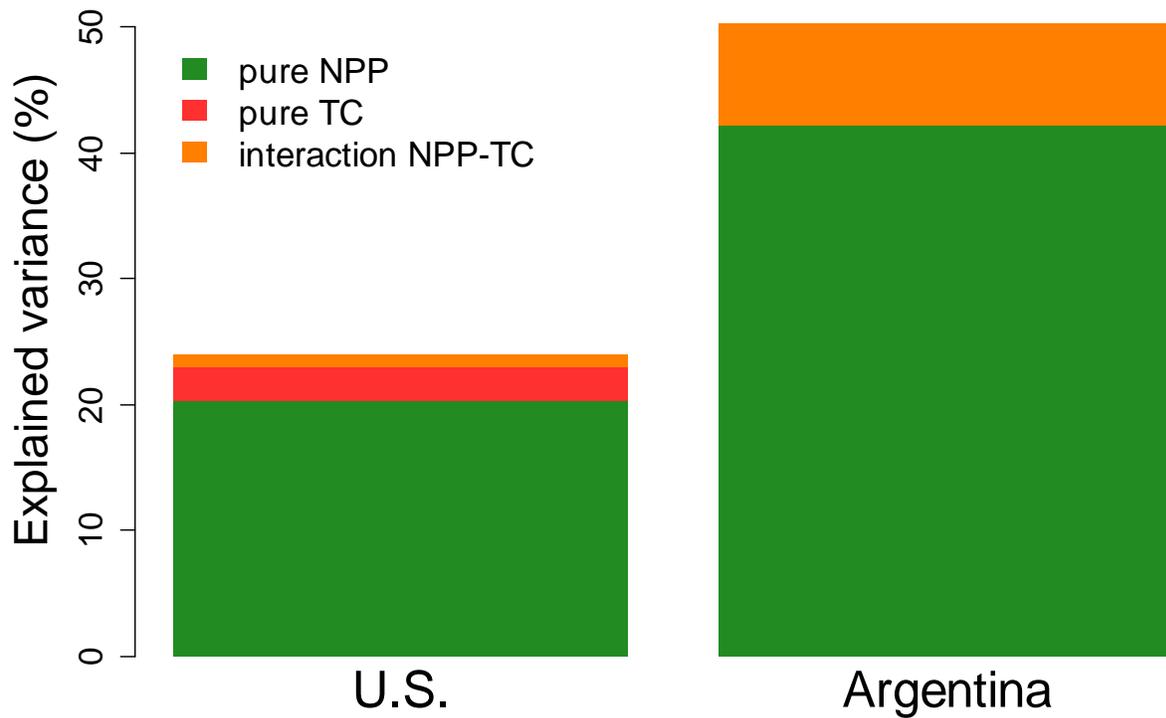
**Figure 2.** Livestock production, net primary productivity (NPP) and tree cover for our study counties. Rangelands not included in the analyses (in grey) are those counties with annual precipitation less than 600 mm or larger than 1260 mm (light gray) or those counties with less than 1000 km<sup>2</sup> in rangelands or less than 25% of their total area in rangelands (dark gray; see Methods).



**Figure 3.** Response models of livestock production to net primary productivity and tree cover in the U.S. and Argentinean rangelands. Equations for response models are shown in Table 1. The red area indicates the NPP range where the impact of tree cover on livestock production is negative. The green area indicates positive effect.  $R_c \text{ km}^{-2}$  = Number of reproductive cows per  $\text{km}^2$ .

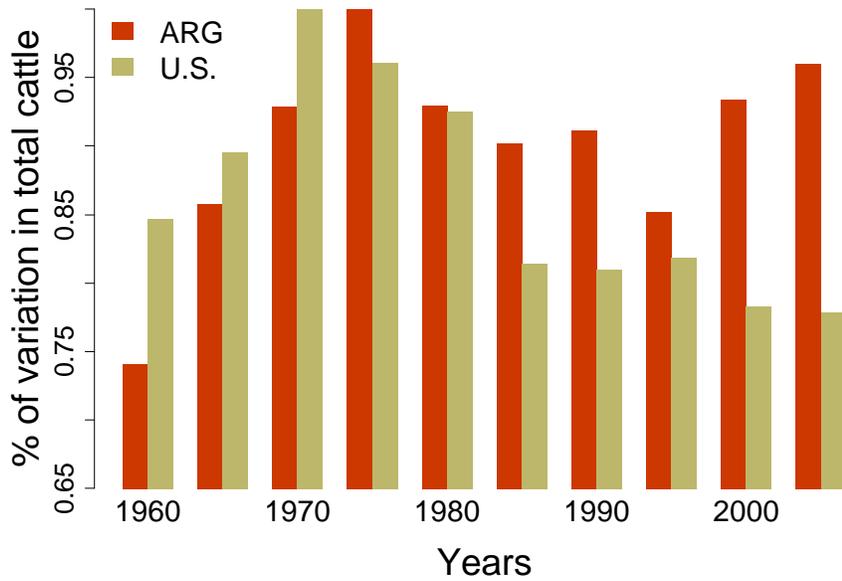
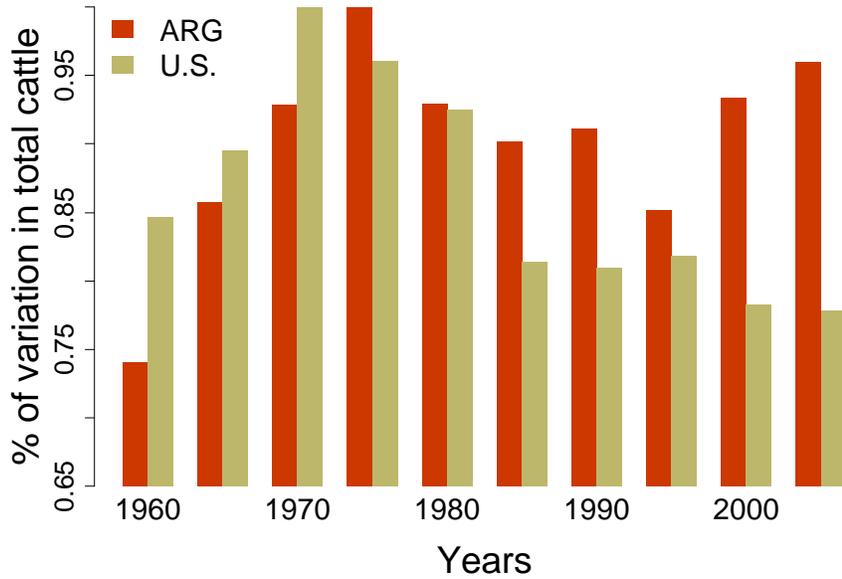


**Figure 4.** Explanatory power of net primary productivity (NPP) and tree cover (TC) on livestock production in the U.S. and Argentinean rangelands as assessed by a variance partitioning analysis. This analysis breaks down the explained variance of the model into a) the pure effects of NPP or TC (i.e. the portion of the variance explained exclusively by one this factors), b) the joint effects of NPP and TC (i.e. the portion of the variance explained jointly by NPP and TC, due to, for example collinearity between them), and c) the interaction between NPP and TC.



## SUPPLEMENTARY INFORMATION

**Figure S1.** Evolution of the number of cattle (top) and agricultural population (bottom) by five-year periods in the U.S. and Argentina. Source: FAOSTAT (<http://faostat.fao.org/>; last accessed Jul 13<sup>th</sup>, 2013).



**Figure S2.** Mean annual precipitation range (600-1260 mm) in relation to tree cover in the U.S. The lower limit of our study area was set at 600 mm, thus excluding those areas where tree cover is marginal (<10%). 1260 mm equals the maximum annual precipitation value for a county in our study area. Annual precipitation from WorldClim database (<http://www.worldclim.com/>), tree cover from MODIS (see Methods section).

